



Significant changes in water pCO₂ caused by turbulence from waterfalls



Zachary William Leibowitz^a, Lorena Aparecida Fortes Brito^a, Placiano Viana De Lima^b, Eneida Maria Eskinazi-Sant'Anna^c, Nathan Oliveira Barros^{d,*}

^a Post-Graduate Program in Ecology of Tropical Biomes, Federal University of Ouro Preto, Campus Morro do Cruzeiro, Ouro Preto – MG, 35400-000, Brazil

^b Post-Graduate Program in Ecology, Federal University of Juiz de Fora, R. José Lourenço Kelmer, Juiz de Fora – MG, 36036-330, Brazil

^c Laboratory of Aquatic Ecology, Evolution and Conservation, Federal University of Ouro Preto, Campus Morro do Cruzeiro, Ouro Preto – MG, 35400-000, Brazil

^d Laboratory of Aquatic Ecology, Federal University of Juiz de Fora, Juiz de Fora – MG, 36036-330, Brazil

ARTICLE INFO

Article history:

Received 15 February 2016

Received in revised form 27 August 2016

Accepted 7 September 2016

Available online 8 October 2016

Keywords:

Carbon dioxide
Outgassing
Waterfalls
River

ABSTRACT

Inland waters are sites of intense carbon processing, stocking and transport. We examined the influence of waterfall-turbulence on CO₂ partial pressures (pCO₂) before and after waterfalls in a tropical river. The results indicated a 51.4% decrease of pCO₂ from up (1375 ± 320 ppm) to downriver (655 ± 58 ppm), suggesting an unaccounted degassing promoted by waterfall-turbulence. This process needs to be better understood in order to more accurately determine the role of freshwater environments in the global carbon balance.

© 2016 Elsevier GmbH. All rights reserved.

Aquatic environments are ecosystems of intense carbon assimilation, processing and exportation (Cole and Caraco, 2001; Duarte and Prairie, 2005). Most of the freshwater ecosystems are CO₂-supersaturated leading to a substantial atmospheric CO₂ efflux (Cole et al., 2007; Tranvik et al., 2009). Some studies have aimed at understanding and evaluating carbon fixation and respiration in aquatic ecosystems, especially in relation to seasonal and annual variations (Raymond et al., 2000; Thomaz et al., 2001); however, studies of the spatial variation of CO₂ in tropical lotic environments are still seldom (Borges et al., 2015a,b).

Determining the CO₂ fluxes at the air-water interface is of fundamental importance in understanding the carbon budget in freshwater ecosystems. Of the criteria that influence gaseous exchange in rivers, the best known are: (1) the boundary layer thickness at the air-water interface and (2) water velocity. These factors change considerably at cascade sites, resulting in three phenomena known as the “jet-flow effect”, “aeration effect” and the “low pressure effect” (Chen et al., 2004). Although there have been some studies examining the release of CO₂ specifically at waterfalls (Herman and Lorah, 1987; Merz-Preiß and Riding, 1999) all have focused on tufa calcification and travertine deposition in karstic

environments. CO₂ spatial variation in lotic systems has already been examined at both the catchment (Schelker et al., 2016) and global-scale (Lauerwald et al., 2015). However, the CO₂ flux from tropical rivers outside of the Amazon and Africa (Borges et al., 2015a,b) have yet to be scrutinized.

Here, we focused on the changes of CO₂ partial pressure (pCO₂) caused by waterfall turbulence. This topic has received attention in Teodoru et al., 2015; as part of a larger study of greenhouse gas dynamics in the Zambezi River in southern Africa. However, as far as we know, the present study is the first to demonstrate and highlight the potential magnitude of changes in pCO₂ caused by waterfall-turbulence in South America.

Sampling was performed on a stretch of the Preto River, a 5th order river located in Rio Preto State Park, in the Cerrado biome of southeastern Brazil (State of Minas Gerais). Samples and *in situ* data were obtained during the dry season for three consecutive days (between 23 and 24 of August 2014). The measurements were all conducted in the morning (between 09:00 and 12:00 AM).

Three falls from Preto River were selected to be sampled based on their high level of turbulence and a long stretch of less turbulent flow leading up to them. Likewise, the study area didn't have any lateral input and the limited macrophyte coverage that was observed is unlikely to cause substantial biogenic changes to pCO₂. Seven random transects (T1 through T7) were performed along the path of the river, and 3 points (right shore, left shore and center)

* Corresponding author.

E-mail address: nathan.barros@ufjf.edu.br (N.O. Barros).

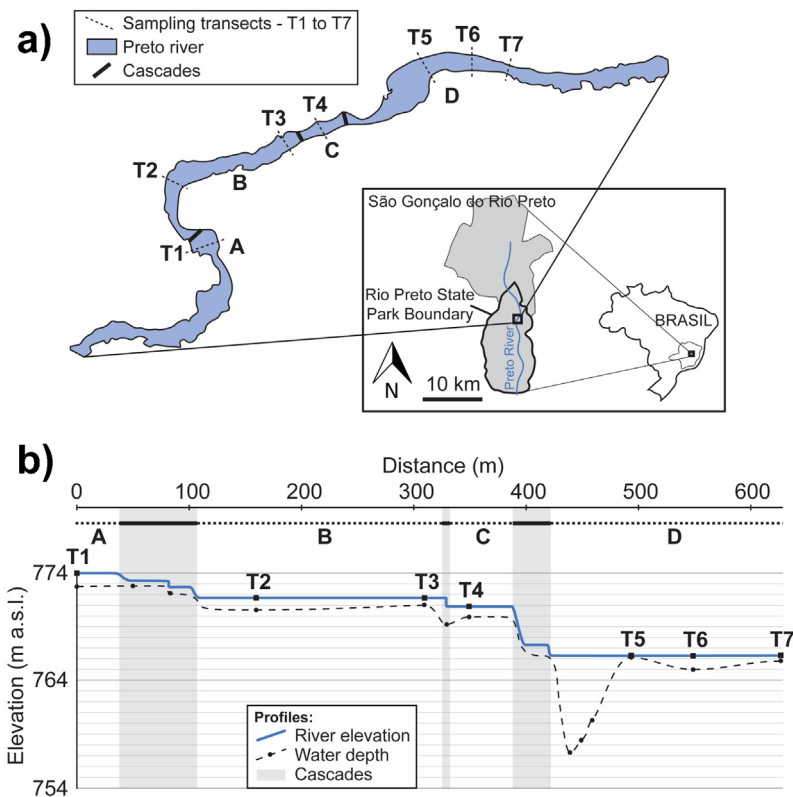


Fig. 1. Map of the section of Preto River analyzed in this study showing both (a) the geographic location and (b) a longitudinal transect of the elevation and water depth changes in meters above sea level. Sampled transects labeled as T1-T7 and regions as A-D.

Table 1
Transect coordinates and their respective regions with limnological results of water temperature (°C), pH, dissolved oxygen (%), alkalinity (uEq/L) and pCO₂ (ppm) as well as river width and average depth along transect.

Region	Transect	Coordinates	Temperature (°C)	pH	Dissolved Oxygen (%)	Alkalinity (uEq/L)	pCO ₂ (ppm)	Width/Average Depth (m)
A	T1	18° 6'43.82"S 43° 20'26.13"W	20.98 ± 0.72	4.63 ± 0.04	104 ± 6.96	6.35	1375 ± 320	35.0/1.3
B	T2	18° 6'41.36"S 43° 20'27.09"W	21.57 ± 0.57	5.13 ± 0.18	109 ± 6.54	3.30	811 ± 269	14.3/1.2
B	T3	18° 6'40.05"S 43° 20'22.85"W	21.63 ± 0.44	5.09 ± 0.05	106 ± 3.43	30.54	971 ± 226	9.5/0.7
C	T4	18° 6'39.59"S 43° 20'21.52"W	21.57 ± 0.12	5.35 ± 0.22	109 ± 4.62	14.40	833 ± 162	6.8/1.0
D	T5	18° 6'37.48"S 43° 20'17.76"W	21.47 ± 0.60	4.23 ± 0.28	109 ± 3.14	11.15	667 ± 111	8.0/0.2
D	T6	18° 6'37.50"S 43° 20'15.92"W	21.17 ± 0.50	5.27 ± 0.27	114 ± 19.64	23.00	696 ± 85	9.1/1.3
D	T7	18° 6'37.85"S 43° 20'14.62"W	20.97 ± 0.63	5.53 ± 0.28	101 ± 19.39	9.00	655 ± 58	9.9/0.5

were sampled at each transect, totaling 21 sampling points per day. This stretch of the river was divided into four regions: i) region A – before the first cascade – T1; region B – after the first cascade – T2 and T3; region C – between the second and third cascade – T4; and region D – after the third cascade – T5, T6 and T7 (Fig. 1).

Limnological parameters measured daily at each sample point with a properly calibrated multiparameter probe (HANNA HI 9829) include: water temperature (°C), dissolved oxygen (%) and pH. Alkalinity was measured once for each point using the volumetric method (APHA, 2005) by titrating the water samples with 0.02 N H₂SO₄. Water velocity was measured with a current meter (AAKER) and used to calculate discharge.

Water samples (30 mL) were collected in triplicate from the surface of each sampling point using a 60 mL syringe and pCO₂ was analyzed using the Headspace Equilibrium Method. The atmospheric air (20 mL) was equilibrated with the river water (30 mL) by vigorously shaking the sample for 3 min, according to Cole and Caraco (1998). Thereafter the gas was transferred to a disposable syringe and the pCO₂ was measured utilizing an infrared gas analyzer (IRGA – Infrared Gas Analyser EGM-4). This was performed 3

times for each sampling point totaling 63 measurements throughout the study period.

Statistical analyses were performed using the Statistical Package for the Social Sciences Software Program – SPSS 20. First a Kolmogorov-Smirnov Normality test was applied; later the homogeneity of the variances was tested using Levene's test. Additionally, an ANOVA One-way was used to account for a difference between regions and between transects. Finally, a Tukey test was used to compare the means. The statistical graphics were generated using the software package JMP, version 11.0 (JMP Statistical Discovery Software, Cary, NC) and the maps were generated using Adobe Illustrator (Adobe Systems, San Jose, CA).

During the study period the Preto River had an average discharge of 3.0 m³/s, was well oxygenated (mean DO% = 106 ± 10), slightly acidic (mean pH = 5.14 ± 0.6), with a mean temperature of 21.5 ± 0.6 °C and a variable, but low alkalinity (16 ± 13 uEq/L; Table 1).

We observed a substantial decreasing trend of pCO₂ from up to downstream of the falls (Fig. 2). A significant difference of pCO₂ between the three points (right shore, left shore and center) within

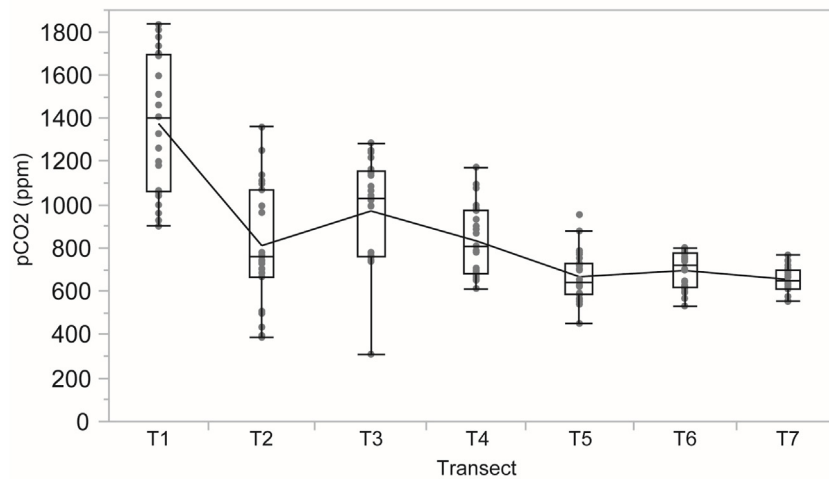


Fig. 2. Box plot of pCO₂ (ppm) values obtained in Preto River with line connecting transect means. Sampling period was from August 23 to 25, 2015.

each sampled transect wasn't found ($F(0.05, 6.226) = 5.96, p = 0.12$). Therefore, here and after we used the average values for the three points to report the pCO₂ in each transect. The values were significantly different (ANOVA, $P < 0.0001$; $df = 3$), between region A (mean = 1375 ± 320 ppm) and regions B, C and D. However, regions B and C (893 ± 259 and 833 ± 162 respectively) were not found to be significantly different from each other, and region D (672 ± 88) was significantly different from the three other regions.

A 41.0% loss of pCO₂ was observed at the first waterfall, while 14.2% and 19.9% losses were observed at the second and third waterfalls, respectively. Considering the total river stretch, a 52.4% decrease in pCO₂ was observed between T1 and T7. The change in pCO₂ associated with CO₂ release upon passage through the waterfalls is the likely cause of the variation in pH and alkalinity. Immediately downstream of the waterfalls (e.g., T2 and T4), alkalinity decreased and pH increased in comparison to the measurements made immediately upstream (Table 1). This is very likely a response to the release of dissolved CO₂, which in turn decreases dissolved inorganic carbon (and hence alkalinity) and increases pH.

The results suggest that the presence of the waterfalls and the distance between the points in this study were important factors in the variability of pCO₂. Considering that T1 lies upstream, far from any disturbances caused by cascades, and that all other transects are located downstream, we can assume that the turbulence caused by the waterfall is an efficient agent to reduce pCO₂ via turbulence-induced degassing. Likewise, long stretches of less turbulent flow after intense turbulence (such as the 150+ m between T2 and T3) may be suitable for small amounts of biogenic respiration and thus slight increases in pCO₂, which could be outgassed at the subsequent waterfalls. The lack of such less turbulent flow between the 2nd and 3rd waterfall may explain the smaller observed change in pCO₂ than expected. In any case, the velocity of the water at waterfalls does not provide a habitat suitable for substantial biogenic changes in CO₂ (fixation or respiration).

Turbulence due to water velocity promotes the breaking of the aqueous boundary layer resistance which, according to Liss (1973), is one of the phenomena that determines the rate of gaseous exchange of slightly soluble gases such as CO₂. Thus, we suggest that the cascades can be responsible for a substantial quantity of CO₂ degassing by promoting turbulence, increasing gas transfer velocities and causing an immediate release of CO₂ to the atmosphere. This can be compared to the 51% of downstream CO₂ release reported for the "turbine" and spillway effect in the Balbina hydroelectric reservoir in the Brazilian Amazon (Kemenes et al., 2011). Our results, however, are markedly lower than the 75% degassing

observed by Teodoru et al., 2015 at the much higher (108 m) and broader (>1 km) Victoria Falls, suggesting that the height and width of waterfalls may play an important role in determining the overall amount of instant CO₂ release.

Despite the recent research on the subject, there are still large uncertainties regarding the factors that regulate CO₂ flux in rivers with waterfalls and the precise quantification of pCO₂ loss at such waterfalls. The loss of carbon by evasion from streams to the atmosphere has been left out of carbon budgets due to the absence of reliable data, although recent studies have indicated that streams and rivers can emit almost 10% of total CO₂ net ecosystem exchange in the United States for example (Butman and Raymond, 2011). In this short communication, we did not present results of CO₂ effluxes (despite having wind and water velocity data) due to the current disagreement of calculating K₆₀₀ (gas transfer velocity standardized to a constant temperature with a Schmidt number of 600) based on wind or water velocity and possibly presenting unrealistic CO₂ efflux values. Despite that, our results showed that waterfalls play a major role in aquatic carbon processing. Therefore, we call attention to the urgent need of better understanding CO₂ degassing from waterfalls, in order to more accurately represent freshwater environments in the global carbon balance.

This could potentially be achieved by measuring pCO₂ before and after a limited number of individual (or groups of small) falls to calculate gas transfer velocities. This data should then use fall height to calculate gas transfer velocity and, combined with data about distance between falls, be used to generate models. Combining these models with technology such as high resolution light detection and ranging (LIDAR) will then yield useful data to be incorporated into larger scale carbon cycle models. While a similar endeavor was undertaken by Hall et al., 2012 to estimate O₂ budgets around rapids in the Colorado River, this has yet to be done for waterfalls, which likely have even higher gas transfer velocities. Realistically, the first studies should focus on larger, highland, tropical rivers that play a potentially larger role on the global carbon balance. However, our results indicate that small rivers in tropical mountains are extremely dynamic and could be degassing significant quantities of CO₂ to the atmosphere over short distances.

Acknowledgments

We would like to thank Mr. André Araújo da Paz, Mathias Schanor, Davi Silva and especially Ms. Capucine Albert for their valuable support throughout the fieldwork period and graphics generation. We also want to thank Rafael Almeida, Martin Schmid and two

anonymous reviewers for their comments that greatly improved the manuscript. Coordinators from postgraduate courses in Ecology from Federal University of Juiz de Fora, Federal University of Ouro Preto and Federal University of Viçosa thank their respective Rectory of Post-Graduate Studies for financial support to the postgraduate ecology field course (ECOMINAS initiative). Graduate fellowships were granted to ZWL and LAFB from the Coordination for the Improvement of Higher Education Personnel (CAPES). EMES is a researcher of the Brazilian National Council for Scientific and Technological Development (CNPq).

References

- APHA, A.P.H.A., 2005. Standard methods for the examination of water and wastewater. Washinton DC, USA.
- Borges, A.V., Darchambeau, F., Teodoru, C.R., Marwick, T.R., Tamooh, F., Geeraert, N., Omengo, F.O., Guérin, F., Lambert, T., Morana, C., Okuku, E., Bouillon, S., 2015a. Globally significant greenhouse-gas emissions from African inland waters. *Nat. Geosci.* 8, 637–642, <http://dx.doi.org/10.1038/ngeo2486>.
- Borges, A.V., Teodoru, C.R., Deborde, J., Vidal, L.O., Lambert, T., Bouillon, S., 2015b. Divergent biophysical controls of aquatic CO₂ and CH₄ in the World's two largest rivers. *Nat. Sci. Rep.* 5, 1–10, <http://dx.doi.org/10.1038/srep15614>.
- Butman, D., Raymond, P. a., 2011. Significant efflux of carbon dioxide from streams and rivers in the United States. *Nat. Geosci.* 4, 839–842, <http://dx.doi.org/10.1038/ngeo1294>.
- Chen, J., Zhang, D.D., Wang, S., Xiao, T., Huang, R., 2004. Factors controlling tufa deposition in natural waters at waterfall sites. *Sediment. Geol.* 166, 353–366, <http://dx.doi.org/10.1016/j.sedgeo.2004.02.003>.
- Cole, J.J., Caraco, N.F., 1998. Atmospheric exchange of carbon dioxide in a low-wind oligotrophic lake measured by the addition of SF₆. *Limnol. Oceanogr.* 43, 647–656, <http://dx.doi.org/10.4319/lo.1998.43.4.0647>.
- Cole, J.J., Caraco, N.F., 2001. Carbon in catchments: connecting terrestrial carbon losses with aquatic metabolism. *Mar. Freshw. Res.*, 101–110, <http://dx.doi.org/10.1071/MF00084>.
- Cole, J.J., Prairie, Y.T., Caraco, N.F., McDowell, W.H., Tranvik, L.J., Striegl, R.G., Duarte, C.M., Kortelainen, P., Downing, J.A., Middelburg, J.J., Melack, J., 2007. Plumbing the global carbon cycle: integrating inland waters into the terrestrial carbon budget. *Ecosystems* 10, 171–184, <http://dx.doi.org/10.1007/s10021-006-9013-8>.
- Duarte, C.M., Prairie, Y.T., 2005. Prevalence of heterotrophy and atmospheric CO₂ emissions from aquatic ecosystems. *Ecosystems*, <http://dx.doi.org/10.1007/s10021-005-0177-4>.
- Hall, R.O., Kennedy, T.A., Rosi-Marshall, E.J., 2012. Air-water oxygen exchange in a large whitewater river. *Limnol. Oceanogr.* Fluids Environ. 2, 1–11, <http://dx.doi.org/10.1215/21573689-1572535>.
- Herman, J.S., Lorah, M.M., 1987. CO₂ outgassing and calcite precipitation in falling spring creek, virginia, U.S.A. *Chem. Geol.* 62, 251–262, [http://dx.doi.org/10.1016/0009-2541\(87\)90090-8](http://dx.doi.org/10.1016/0009-2541(87)90090-8).
- Kemenes, A., Forsberg, B.R., Melack, J.M., 2011. CO₂ emissions from a tropical hydroelectric reservoir (Balbina, Brazil). *J. Geophys. Res. Biogeosci.* 116, 1–11, <http://dx.doi.org/10.1029/2010JG001465>.
- Lauerwald, R., Laruelle, G.G., Hartmann, J., Ciais, P., Regnier, P. a. G., 2015. Spatial patterns in CO₂ evasion from the global river network. *Glob. Biogeochem. Cycles* 29, 534–554, <http://dx.doi.org/10.1002/2014GB004941>.
- Liss, P.S., 1973. Processes of gas exchange across an air-water interface. *Deep. Res. Oceanogr. Abstr.* 20, 221–238, [http://dx.doi.org/10.1016/0011-7471\(73\)90013-2](http://dx.doi.org/10.1016/0011-7471(73)90013-2).
- Merz-Preiß, M., Riding, R., 1999. Cyanobacterial tufa calcification in two freshwater streams: ambient environment, chemical thresholds and biological processes. *Sediment. Geol.* 126, 103–124, [http://dx.doi.org/10.1016/S0037-0738\(99\)00035-4](http://dx.doi.org/10.1016/S0037-0738(99)00035-4).
- Raymond, P.A., Bauer, J.E., Cole, J.J., 2000. Atmospheric CO₂ evasion, dissolved inorganic carbon production, and net heterotrophy in the York River estuary. *Limnol. Oceanogr.* 45, 1707–1717, <http://dx.doi.org/10.4319/lo.2000.45.8.1707>.
- Schelker, J., Singer, G.A., Ulseth, A.J., Hengsberger, S., Battin, T.J., 2016. CO₂ evasion from a steep, high gradient stream network: importance of seasonal and diurnal variation in aquatic pCO₂ and gas transfer. *Limnol. Oceanogr.* 2, <http://dx.doi.org/10.1002/lno.10339>.
- Teodoru, C.R., Nyoni, F.C., Borges, A.V., Darchambeau, F., Nyambe, I., Bouillon, S., 2015. Dynamics of greenhouse gases (CO₂, CH₄ N₂O) along the Zambezi River and major tributaries, and their importance in the riverine carbon budget. *Biogeosciences* 12, 2431–2453, <http://dx.doi.org/10.5194/bg-12-2431-2015>.
- Thomaz, S.M., Enrich-Prast, A., Goncalves, J.J.F., Dos Santos, A.M., Esteves, F.A., 2001. Metabolism and gaseous exchanges in two coastal lagoons from rio de janeiro with distinct limnological characteristics. *Braz. Arch. Biol. Technol.* 44, 433–438, <http://dx.doi.org/10.1590/S1516-89132001000400015>.
- Tranvik, L.J., Downing, J. a., Cotner, J.B., Loiselle, S. a., Striegl, R.G., Ballatore, T.J., Dillon, P., Finlay, K., Fortino, K., Knoll, L.B., 2009. Lakes and reservoirs as regulators of carbon cycling and climate. *Limnol. Oceanogr.* 54, 2298–2314, <http://dx.doi.org/10.4319/lo.2009.54.6.part.2.2298>.